

Mycotoxins and human health

Summary

Mycotoxins have been investigated in relation to a wide range of adverse human health effects, but the evidence for all but a small number of associations is limited. Thus, the full impact on human health of the widespread exposure to mycotoxins remains to be defined. The main exception is for aflatoxins; epidemiological, experimental, and mechanistic studies have contributed to establishing aflatoxins as a cause of human liver cancer, with a particularly elevated risk in people chronically infected with hepatitis B virus. In addition, acute aflatoxicosis after exposure to high dietary toxin levels has been demonstrated. The impairment of child growth by aflatoxin exposure early in life remains an important subject of study. More information is also required on the potential immune effects of aflatoxins,

especially in vulnerable populations. For fumonisins, studies indicate a possible role in oesophageal cancer and in neural tube defects, although no definitive conclusions can be drawn at present. For deoxynivalenol and other trichothecenes, exposure has been linked to acute poisoning outbreaks in large numbers of subjects. For ochratoxin A and zearalenone, the human health effects remain undefined. The limited tools available to accurately assess human exposure to mycotoxins and the relative paucity of epidemiological studies need to be addressed if the full extent of the adverse effects of these common dietary contaminants is to be understood and adequate public health measures taken. In this respect, newly established biomarkers of exposure at the individual level are proving valuable in improving exposure assessment in epidemiological studies.

1. Introduction

This chapter covers the effects on human health of the major mycotoxins occurring in foods. This chapter also includes information on mechanisms of action of mycotoxins in humans where relevant to the adverse health effects under consideration. No attempt is made at a comprehensive review, but at appropriate points we refer to more extensive accounts.

The major source of human exposure to mycotoxins is consumption of contaminated foods. Exposure is highest when those foods are dietary staples, such as maize, groundnuts, or various other cereals. Exposures to metabolites or parent toxins may also occur by consumption of contaminated milk and milk products. We covered the dietary sources of mycotoxin exposure in some detail in Chapter 1. In this chapter, we also consider

occupational exposures in granaries and other food and feed processing due to mycotoxins contained in dusts from contaminated grains. The specific effects due to particular mycotoxins are discussed in Sections 2–6. Additional, more general information on health problems associated with mycotoxins in grain dusts is covered in Section 7.

The human health effects considered here encompass acute poisoning, cancer, other chronic diseases, and biological effects, including growth impairment and immunomodulation.

One of the major limitations in assessing the effects of mycotoxins on health has been the inability to accurately assess exposure at the individual level. The development of validated biomarkers for aflatoxins has greatly assisted epidemiological studies and allowed an evaluation of aflatoxins in relation to cancer, aflatoxicosis, child growth impairment, and immune effects (see Wild and Gong, 2010). Development of validated biomarkers for fumonisins (Wild and Gong, 2010; Van der Westhuizen *et al.*, 2011) and deoxynivalenol (Meky *et al.*, 2003; Turner *et al.*, 2008a, 2008b, 2008c) also offers promise for future studies of the human health effects of these mycotoxins. However, the biomarker field for mycotoxins also offers a cautionary tale: an unvalidated biomarker for ochratoxin A in plasma or serum has been used to assess dietary exposure to this toxin, but subsequent careful duplicate diet studies have shown that this biomarker does not reflect intake at the individual level (Gilbert *et al.*, 2001). Nevertheless, the availability of biomarkers to measure exposure to mycotoxins provides new opportunities for more systematic monitoring of exposure in populations as well as improved etiological studies.

2. Aflatoxins

Aflatoxins are produced in a wide range of commodities by *Aspergillus flavus* and *A. parasiticus*, and occasionally other *Aspergillus* species. The commodities most at risk are maize and groundnuts in tropical areas (see Chapter 1).

2.1 Mechanisms

Until recently, attention on aflatoxins has been focused on their carcinogenic effects. For more detailed information, see the extensive reviews in IARC (2002), WHO (2002), and Wild and Gong (2010). Consideration is usually given to the naturally occurring aflatoxins in the diet – aflatoxins B₁ (AFB₁), B₂ (AFB₂), G₁ (AFG₁), and G₂ (AFG₂) – or to AFB₁ alone, or more rarely to aflatoxins M₁ (AFM₁) and M₂ (AFM₂), the hydroxylation products of AFB₁ and AFB₂, respectively, that occur in milk. This distinction between the type of aflatoxin exposures under consideration is important but is rarely considered, particularly when studying adverse health effects other than cancer.

Given the focus on mutagenicity and carcinogenicity, most studies have been of AFB₁, which, due to the presence of a double bond at the 8,9 position, can be metabolized to the reactive AFB₁-8,9-epoxide, which binds to cellular macromolecules including DNA (for more detail, see Wild and Turner, 2002). The major DNA adduct is AFB₁-N7-guanine, and this pro-mutagenic lesion commonly results in a G → T transversion mutation. AFB₁-N7-guanine can also be detected in the urine and used as an exposure biomarker in epidemiological studies. AFB₂ and AFG₂ are generally considered to be far less biologically active due to the absence of the 8,9 double bond. AFG₁ can be bioactivated to the 8,9-epoxide

but is less mutagenic than AFB₁, reflecting the steric chemistry of the respective epoxides; the AFB₁-8,9-epoxide intercalates more readily into the DNA double helix than does the equivalent AFG₁ molecule, resulting in higher levels of DNA adduct formation for a given dose. Minimal information exists about the importance of the reactive epoxide for the non-mutagenic actions of aflatoxins or indeed about the biological effects of aflatoxins independent of metabolic activation to the 8,9-epoxide. Unlike AFM₂, AFM₁ contains an 8,9 double bond and hence can be bioactivated to the reactive 8,9-epoxide.

A significant observation in terms of aflatoxin carcinogenicity is the association between exposure and a specific mutation in the *TP53* tumour suppressor gene in liver cancer (hepatocellular carcinoma [HCC]). In HCC tumours from patients who are from regions endemic for aflatoxin and who are chronically infected with hepatitis B virus (HBV), a high prevalence exists of a specific missense mutation in the gene, namely an AGG → AGT (Arg → Ser) point mutation at codon 249 (codon 249^{ser}) (IARC, 2002; Hussain *et al.*, 2007). This mutation is extremely rare in HCC associated with HBV in areas where aflatoxins are uncommon, but it is as yet unclear whether HBV infection influences occurrence of the mutation in HCC from aflatoxin-endemic areas.

The major human cytochrome P450 (CYP) enzymes involved in aflatoxin metabolism are CYP3A4, 3A5, and 1A2, and the predominant site of bioactivation is the liver, although CYP3A4 expression in the human intestine means that metabolism may also occur in that organ (Wild and Turner, 2002; Kamdem *et al.*, 2006; Thelen and Dressman, 2009). The contribution of these enzymes to AFB₁ metabolism in exposed people will depend on both

the affinity and level of expression of the different enzymes; CYP3A4 appears to be the most important in generating the *exo*-8,9-epoxide, and the relative contribution of CYP3A5, which also produces the *exo*-8,9-epoxide, varies by individual (Kamdem *et al.*, 2006). In fact, CYP3A5 expression is polymorphic and varies by ethnic group; for example, 40% of African Americans show no expression due to identified genetic polymorphisms. Such polymorphisms may affect sensitivity to the toxic effects of aflatoxins (Wojnowski *et al.*, 2004). CYP1A2 leads predominantly to formation of the hydroxylated AFM₁ metabolite and the AFB₁-*endo*-8,9-epoxide, which does not form DNA adducts.

Given the fact that aflatoxin is known to cross the placenta, it is also of interest that CYP3A7, a major CYP in human fetal liver, can activate AFB₁ to the 8,9-epoxide (Kamataki *et al.*, 1995; Wild and Turner, 2002). Indeed, aflatoxin adducts have been identified in cord blood (Wild *et al.*, 1991; Turner *et al.*, 2007), indicating that environmental levels of aflatoxin are bioactivated to the reactive metabolites in utero.

Detoxification of the aflatoxin *exo*- and *endo*-epoxides occurs mainly through glutathione S-transferase-mediated conjugation to reduced glutathione (Guengerich *et al.*, 1998). The epoxides can also undergo rapid non-enzymatic hydrolysis to AFB₁-8,9-dihydrodiol, which in turn forms a dialdehyde phenolate ion with an opened ring. The dihydrodiol can react with the ϵ -amino group of lysine in serum albumin to form aflatoxin–albumin adducts, which are often used as exposure biomarkers (Wild and Gong, 2010). In a further metabolic step, aflatoxin aldehyde reductase catalyses the NADPH-dependent reduction of the dialdehyde phenolate ion to a dialcohol (Johnson *et al.*, 2008).

An understanding of the metabolism, DNA damage, and induction of mutations in people exposed to aflatoxins in the diet has contributed to the overall assessment of their adverse health effects (Groopman *et al.*, 2008; Wild and Gong, 2010). The major health effects linked to aflatoxin exposure are described briefly here.

2.2 Aflatoxicosis

Sporadic historical accounts of human poisoning with aflatoxins were reported, but these early studies were not definitive in assigning causation (Hall and Wild, 1994). In 1974, hepatitis cases due to aflatoxicosis in Western India (Krishnamachari *et al.*, 1975) were associated with consumption of maize contaminated with *A. flavus*. Patients exhibited jaundice preceded by fever, vomiting, and anorexia, with subsequent progression to ascites and oedema in lower extremities. In maize from households where cases occurred, the aflatoxin levels were extremely high, 6.25–15.6 mg/kg, and the estimated daily ingestion of aflatoxins was 2–6 mg in adults.

In Kenya in 1981, another outbreak of acute hepatitis was associated with aflatoxin poisoning (Ngindu *et al.*, 1982). Patients were diagnosed with jaundice preceded by abdominal discomfort, anorexia, general malaise, and low-grade fever; tachycardia and oedema were also observed. Maize from two affected households contained up to 3.2 mg/kg and 12 mg/kg AFB₁. An additional report came from an incident in Malaysia in 1988, where 13 children died from acute hepatic encephalopathy after consuming noodles (Lye *et al.*, 1995).

In 2004, well-documented cases of aflatoxicosis occurred in Kenya, close to the locality of the cases reported in 1981 (Azziz-

Baumgartner *et al.*, 2005; Lewis *et al.*, 2005). These outbreaks resulted in several hundred deaths associated with consumption of maize heavily contaminated with aflatoxin. A case–control study of aflatoxicosis, defined as acute jaundice of unknown origin, found that aflatoxin levels in foods from affected households were much higher than those in foods from unaffected households. Similar differences between cases and controls were found when aflatoxin biomarker levels in blood were examined (Azziz-Baumgartner *et al.*, 2005; McCoy *et al.*, 2008).

The association of aflatoxin contamination of maize with acute hepatitis and aflatoxicosis is well supported by the evidence, most notably by the observations in Kenya. It is of interest that aflatoxicosis has been reported only in communities where maize is the dietary staple. This reflects both high levels of aflatoxins in maize and high daily intakes (300–500 g) of this staple commodity. In addition, however, the role of co-contaminating mycotoxins, notably fumonisins, has not been assessed, and these may contribute to the acute toxicity observed.

The level of aflatoxin intake associated with aflatoxicosis and death has been estimated (Wild and Gong, 2010). The intake of total aflatoxins estimated to result in a risk of fatality was > 1 mg/day, i.e. > 20 μ g/kg body weight (bw)/day in adults. It was considered that aflatoxicosis without fatality may occur with 5–10-fold lower doses. Further estimates suggested that the total intake of AFB₁ associated with half the exposed people dying (i.e. the median lethal dose [LD₅₀]) would be 0.54–1.62 mg/kg bw, a similar magnitude to the LD₅₀ value reported for rabbits, cats, dogs, pigs, and baboons (Wild and Gong, 2010). Therefore, daily exposure to staple foods consumed at several hundred

grams per day and contaminated with $\geq 5000 \mu\text{g}/\text{kg}$ of aflatoxins may lead to death in humans. Daily consumption of foods with $> 1000 \mu\text{g}/\text{kg}$ may lead to aflatoxicosis.

It is of great concern that these contamination levels in maize that are associated with aflatoxicosis and death are only 10–100 times the levels that occur regularly in many parts of sub-Saharan Africa. Despite the demonstration that heavy contamination of maize with aflatoxins does lead to aflatoxicosis and death, these outbreaks continue to occur. Thus, in affected parts of the world, there is an urgent need for a rapid field test that, as part of a preventive strategy, can detect dangerously high levels (e.g. $> 1000 \mu\text{g}/\text{kg}$) of aflatoxins in cereals and nuts, as well as an emergency response analogous to those in place for outbreaks of infectious diseases.

2.3 Liver cancer

The International Agency for Research on Cancer (IARC) has classified naturally occurring mixtures of aflatoxins as Group 1, carcinogenic to humans (IARC, 2002) (Table 6.1). Before the 1990s, most studies of aflatoxins and HCC were either ecological correlation studies or case–control studies. The ecological studies did not always consider infection with HBV and hepatitis C virus, had relatively crude estimates of aflatoxin intakes, and had limitations in diagnosis and registration of HCC cases. Despite this, most showed a positive correlation between estimated aflatoxin intakes and HCC rates in a given region (IARC, 1993, 2002).

Prospective cohort studies of improved design, begun in the late 1980s and 1990s in South-East Asia, used biomarkers of exposure to both HBV and aflatoxins. These studies provided strong evidence of a more than multiplicative interaction

between these two factors in relation to increased HCC risk (Qian *et al.*, 1994; Wang *et al.*, 1996; IARC, 2002). In a more recent follow-up of the cohort in Taiwan, China, Wu *et al.* (2009) conducted the largest nested case–control study to date and reported that the combined effect of AFB₁ exposure and HBV infection was more consistent with an additive model than with the multiplicative one observed in the original report of Wang *et al.* (1996). However, after examining the plasma of HCC patients, cirrhosis patients, and controls for the *TP53* gene mutation (codon 249^{ser}; AGG → AGT), Kirk *et al.* (2005) showed that the increased risk associated with the presence of both the 249^{ser} mutation and HBV infection was consistent with a multiplicative effect of exposure to aflatoxin and chronic HBV infection.

The HCC risk from exposure to aflatoxins in the absence of chronic HBV infection is difficult to assess in populations where HBV infection is widespread. A review by Omer *et al.* (2004) reported 1.7–3.4-fold increased risks in individuals exposed to aflatoxins without chronic HBV infection. Wu *et al.* (2009) reported a similar magnitude of increased HCC risk in subjects positive only for aflatoxin exposure biomarkers, but occult HBV infections in some of the individuals in these studies cannot be ruled out.

The overall evidence from epidemiological studies shows a particularly elevated risk of HCC from aflatoxin exposure in individuals chronically infected with HBV and reasonable evidence that an increased risk also exists in individuals exposed to aflatoxins without chronic HBV infection. Given that > 350 million chronic HBV carriers exist worldwide, many living in aflatoxin-endemic areas, the need to reduce aflatoxin exposure remains highly relevant for cancer prevention.

2.4 Cirrhosis

To date, little information exists on the risk of liver cirrhosis in relation to aflatoxin exposure. A case–control study in The Gambia (Kuniholm *et al.*, 2008) reported that increasing lifetime groundnut intake (a surrogate for aflatoxin consumption) was associated with a significantly increased risk of cirrhosis, approaching 3-fold with the highest level of consumption. The presence of the codon 249^{ser} mutation associated with aflatoxin was also associated with a similar magnitude of increased risk of cirrhosis. However, further studies are needed before any conclusions can be drawn about aflatoxin and cirrhosis. This is an area that merits more attention, given the large burden of cirrhosis worldwide.

2.5 Immune effects

The immunomodulatory effects of aflatoxins have been considered in experimental studies in cell models and animals as well as in observations of farm animals (IARC, 1993, 2002; WHO, 2002; Williams *et al.*, 2004). However, only a few studies have considered the association between aflatoxin exposure and immune parameters in human populations. Two such studies have been reported from The Gambia (Allen *et al.*, 1992; Turner *et al.*, 2003). The first provided some evidence that children with higher aflatoxin exposure were more likely to have malaria parasitaemia, but no significant associations were observed with experience of malaria infection, antibody titre to asexual stages of *Plasmodium falciparum*, or lymphoproliferative responses. The second study investigated the effect of aflatoxin exposure on cell-mediated immunity (skin test), antibody titres (in response to rabies and pneumococcal vaccines), and salivary immunoglobulin A (IgA). No associations were found between

aflatoxin exposure and either the skin test or antibody titres, but higher aflatoxin exposure was associated with lower salivary IgA, suggesting that aflatoxin exposure could modulate mucosal immunity.

From Ghana, two studies have been reported that compared aflatoxin biomarker levels and subsets of peripheral blood cells in adults (Jiang *et al.*, 2005, 2008). In the first, a higher aflatoxin biomarker level was associated with a lower percentage of CD3+ and CD19+ cells (B lymphocyte antigens) expressing the CD69+ activation marker and with lower percentages of CD8+ T cells expressing perforin and granzyme A. In the second study, a higher aflatoxin biomarker level was associated with lower percentages of CD8+ cells expressing perforin and of CD19+ cells expressing CD69+. In addition, HIV-positive individuals with higher aflatoxin biomarker levels had significantly lower percentages of CD4+ T regulatory cells and naive CD4+ T cells compared with HIV-positive individuals with lower aflatoxin biomarker levels.

Overall, the studies of immunomodulation do not permit conclusions to be drawn about the impact of environmental levels of aflatoxin exposure on human immunity and susceptibility to infectious disease. Nevertheless, the data suggest that immune parameters could be affected in populations exposed chronically to aflatoxins. If this were to be proven, the impact would add greatly to the burden of disease related to cancer and aflatoxicosis.

2.6 Child growth impairment

Children are chronically exposed to high levels of aflatoxins in areas where food contamination is endemic. Exposure begins in utero and continues throughout early life,

although the breastfeeding period provides some respite from high daily intakes. Studies in several animal species indicate that aflatoxin exposure can severely affect growth and development. However, until recently such effects had not been considered in human populations.

Early studies explored the link between aflatoxin exposure and kwashiorkor (Hendrickse *et al.*, 1982), but no firm conclusions could be drawn due to various weaknesses in study design (Hall and Wild, 1994). A study in rural Kenya in the 1980s linked aflatoxin detection in mothers' blood with significantly lower birth weights of female babies (De Vries *et al.*, 1989). A more recent study in Kisumu District, Kenya, showed a significantly greater prevalence of wasting (low weight for height) in children fed cereals with high aflatoxin contamination, compared with those whose cereals had lower aflatoxin levels (Okoth and Ohingo, 2004).

A series of studies has been conducted in West African children exposed to aflatoxins early in life. In the first of these, a cross-sectional study of children aged 1–5 years in Benin and Togo, a striking inverse association was found between aflatoxin–albumin adduct level and growth (Gong *et al.*, 2002). In a subsequent 8-month longitudinal study, a strong negative correlation was observed between aflatoxin–albumin adduct level and height increase (Gong *et al.*, 2004). The highest quartile of aflatoxin–albumin adducts was associated with a mean reduction of 1.7 cm in height increase compared with the lowest quartile. Recently, an association was also found between exposure to aflatoxin in utero and impaired growth during the first year of life in children in The Gambia (Turner *et al.*, 2007). This finding suggests that the consumption of aflatoxin-contaminated food during pregnancy may have effects on the child after birth.

In summary, growth faltering in West African children may occur at the time of introduction of solid foods, when high exposure to aflatoxin occurs. The dose–response relationships between aflatoxin biomarker levels and growth effects are also consistent with a causal effect. However, at this time other confounding factors cannot be excluded as explanations for these associations. The mechanisms of action by which aflatoxin may exert an effect on growth are currently unknown, although the possibility of a compromised intestinal integrity, through altered barrier function as a consequence of endothelial cell toxicity or immune suppression, is a valid hypothesis that should be explored further (Wild and Gong, 2010).

In areas where aflatoxin is common, namely sub-Saharan Africa and South Asia, 7.1 million children died under the age of 5 years in 2008. It is estimated that about 50% (3.55 million) of these deaths are related to undernutrition and poor growth (Black *et al.*, 2003). If aflatoxin exposure were to be responsible for even a few per cent of these deaths, the total number would be tens of thousands per year.

2.7 Occupational exposures

AFB₁ concentrations of up to 612 µg/kg have been reported in airborne dusts during the handling of contaminated maize and groundnuts (Miller, 1994a; Sorenson, 1999). Most aflatoxin was contained in the < 7 µm and 7–11 µm particle size ranges. In grain dusts, a substantial fraction of the aflatoxin is contained in the spores of *A. flavus* and *A. parasiticus* (Miller, 1994a, 1994b).

Retrospective studies of feed processing workers in Denmark reported elevated risks of HCC, gall bladder cancer, and extrahepatic bile duct cancer in this population exposed occupationally, with a

2-3-fold increased risk after a 10-year latency period. Inhalation exposure to aflatoxin (170 ng/day) was reported to be the most likely explanation. Some evidence of elevated aflatoxin biomarker levels related to the handling of contaminated feeds was also reported (Olsen *et al.*, 1988; Autrup *et al.*, 1991, 1993). A risk assessment model suggested that exposure to AFB₁ in airborne dust may pose little significant risk during maize harvest and elevator loading/unloading but a relatively high risk during swine feeding and storage bin cleaning (Liao and Chen, 2005).

3. Fumonisin

Fumonisin occurs in maize and, much less commonly, other cereals, as a result of infection with *Fusarium verticillioides* and related species (see Chapter 1). An Environmental Health Criteria document for fumonisin B₁ (FB₁) has been published (WHO, 2000a), and the Joint FAO/WHO Expert Committee on Food Additives (JECFA) has established a provisional maximum tolerable daily intake (PMTDI) of 2 µg/kg bw/day for FB₁, FB₂, and FB₃ alone or in combination (WHO, 2001, 2002). This PMTDI is based not on tumorigenicity data but on the no-observed-effect level (NOEL) for nephrotoxicity in rodents of 0.2 mg/kg bw/day, divided by a safety factor of 100 (WHO, 2001).

3.1 Mechanisms

Fumonisin may exert their biological effects through several different mechanisms. FB₁ genotoxicity is somewhat unclear; negative results were obtained from several genotoxicity assays, but other *in vitro* studies reported that FB₁ induced micronuclei and chromosomal aberrations (Ehrlich *et al.*, 2002; IARC, 2002). The DNA damage may be a result of stimulation of oxidative

damage and lipid peroxidation (Stockmann-Juvala and Savolainen, 2008). This finding is consistent with increased oxidative DNA damage and malondialdehyde adducts in rat liver and kidney (Domijan *et al.*, 2006) and lipid peroxidation (Abel and Gelderblom, 1998) *in vivo* after FB₁ treatment. Fumonisin-induced carcinogenesis in the liver proceeds through initiation and promotion stages in a manner similar to that for genotoxins and is dependent on the dose and time of exposure (Gelderblom *et al.*, 2008a). Nevertheless, no evidence has been found for direct interaction of fumonisin with DNA nor for its metabolism to a reactive metabolite (WHO, 2001; IARC, 2002).

FB₁ disrupts *de novo* sphingolipid biosynthesis by inhibition of the enzyme ceramide synthase (Merrill *et al.*, 2001), resulting in many effects on signalling pathways and cell functions that are dependent on ceramide, sphingoid bases, sphingoid base 1-phosphates, and complex sphingolipids (Dragan *et al.*, 2001; Merrill *et al.*, 2001). These include effects on apoptosis and mitosis, thus potentially contributing to carcinogenesis through an altered balance of cell death and replication (Stockmann-Juvala and Savolainen, 2008). Disruption of sphingolipid metabolism leads to changes in the sphinganine-to-sphingosine ratio, with increased sphinganine tissue concentrations, which correlate closely with the *in vivo* toxicity and carcinogenicity of fumonisin (Riley *et al.*, 2001). Such changes were demonstrated in rat liver and mouse kidney at carcinogenic doses of FB₁ (Voss *et al.*, 2002). Disruption of cholesterol, phospholipid, and fatty acid synthesis and interaction with ceramide have been proposed to play key roles in the differential growth patterns of altered hepatocytes during cancer promotion in the liver (Gelderblom *et al.*, 2008b).

The role of fumonisin in immunomodulation has also been highlighted through changes in cytokine levels *in vitro* and *in vivo* in animal models (Sharma *et al.*, 2000) and effects on antibody vaccine responses in pigs exposed to FB₁ (Taranu *et al.*, 2005; Stockmann-Juvala and Savolainen, 2008).

Animal experiments have shown that fumonisin is rapidly excreted unmetabolized from the gut (Shephard *et al.*, 1994a, 1994b; Martinez-Larranaga *et al.*, 1999). It is not known whether the gut microflora metabolize fumonisin, although hydrolysed FB₁ has been detected in faeces of vervet monkeys (Shephard *et al.*, 1994a, 1994b), pigs (Fodor *et al.*, 2008), and ruminants (WHO, 2000b). Studies using radiolabelled fumonisin failed to show any metabolism by primary hepatocytes or hepatic esterases and lipases (Cawood *et al.*, 1994).

Analyses of human faeces has revealed unmetabolized FB₁ and FB₂ (Chelule *et al.*, 2000, 2001). The presence of FB₁, FB₂, and FB₃ in human hair has been demonstrated, suggesting that fumonisin is absorbed from the gut after ingestion of contaminated maize (Sewram *et al.*, 2003). More recently, FB₁ in urine has been reported in individuals consuming large amounts of maize (Gong *et al.*, 2008). Direct evidence therefore exists that human populations are exposed to fumonisin after absorption, although the actual levels and the possible risk this poses still need to be quantified.

Probable daily intake values of fumonisin, determined using a validated dietary assessment tool, showed that intakes of up to 10 times the PMTDI can occur in individuals in a rural, subsistence farming community where maize is the main dietary staple (Burger *et al.*, 2010). Drinking home-brewed maize beer further increased the level of exposure. Estimates of daily

fumonisin intake in rural communities in Guatemala (Torres *et al.*, 2007) have also shown that total fumonisin intake can potentially be > 10 times the recommended PMTDI.

3.2 Acute poisoning

No confirmed cases have been reported of acute human poisoning due to fumonisin exposure. Part of the difficulty in discerning a specific effect of fumonisins in acute poisoning is their co-occurrence with other mycotoxins, notably aflatoxins and trichothecenes. For example, one poisoning outbreak in India occurred where fumonisin contamination of foods was reported, notably in unleavened bread prepared from mouldy sorghum or maize, and where symptoms were characterized by abdominal pain, borborygmi, and diarrhoea (Bhat *et al.*, 1997). However, assays of other mycotoxins potentially present were not reported.

An outbreak of human intoxication related to the consumption of maize gruel prepared from mouldy maize powder occurred in Guangxi Province, China, in 1989 (Li *et al.*, 1999). In this case, co-occurrence with trichothecenes was found, and based on an average maize meal intake of 200 g/person (60 kg bw) per day, the total daily dietary intake of deoxynivalenol and FB₁ was 80 µg/kg bw and 2.3 µg/kg bw, respectively.

These incidents highlight the need for comprehensive analyses of mycotoxins in contaminated foods and in biological samples when acute poisoning outbreaks occur.

3.3 Cancer

Ecological studies in the former Transkei region of South Africa showed that both *F. verticillioides* and fumonisin contamination of maize were positively correlated with oesophageal cancer incidence rates (Marasas, 2001; IARC, 2002).

Similar correlations have been reported in China (Sun *et al.*, 2007). Other reports have associated maize consumption per se with high oesophageal cancer incidence rates but did not consider fumonisin exposure specifically (Franceschi *et al.*, 1990). In these studies, other fungi and their mycotoxins were generally also present, and to date no analytical studies have been conducted that specifically link FB₁ to human cancer at any organ site (IARC, 1993, 2002; WHO, 2001).

Based on the disruption of sphingolipid biosynthesis mentioned above, the serum sphinganine-to-sphingosine ratio was used as an exposure biomarker in a nested case-control study of oesophageal cancer in China (Abnet *et al.*, 2001), but no association was found between biomarker levels and cancer risk. One study conducted in China did report that the sphinganine-to-sphingosine ratio in urine was significantly increased in males estimated to have consumed > 110 µg/kg bw/day of FB₁ (Qiu and Liu, 2001). However, in all subsequent reports, no association was observed between sphingoid bases or sphinganine-to-sphingosine ratios in the plasma and urine and individual fumonisin exposure, suggesting that these biomarkers are not sufficiently sensitive for monitoring exposure in human populations (Nikiéma *et al.*, 2004; van der Westhuizen *et al.*, 2010; Xu *et al.*, 2010).

Many risk factors exist for the development of oesophageal cancer. These differ among geographical regions with respect to demography, ethnicity, genetic susceptibility, cultural practices, and socio-economic and nutritional status. The use of home-grown maize as one of the main dietary staples, coupled with an underlying poor socioeconomic status, could implicate fumonisins as a contributing factor in the

development of oesophageal cancer. However, confounding by other risk factors and the possible interactions of fumonisins with other mycotoxins should be considered in future studies.

Some experimental studies have reported a synergistic interaction between AFB₁ and FB₁ in the development of liver cancer (Carlson *et al.*, 2001; Gelderblom *et al.*, 2002). Perhaps due to the focus on oesophageal cancer, the role of fumonisins in cancer in other organs has been largely unexplored. However, given the interactions found experimentally, the co-contamination of crops by aflatoxins and fumonisins, and the fact that both toxins occur in populations with a high prevalence of HBV infection, a role for fumonisins in HCC is plausible. Some ecological correlation studies provide support for this hypothesis (Ueno *et al.*, 1997; Li *et al.*, 2001; Sun *et al.*, 2007). The possible interaction between various mycotoxins (fumonisins, aflatoxins, trichothecenes) and the microcystins B (algal toxins) in the development of HCC merits more investigation. Interactions of fumonisins with different dietary constituents could also have an impact on the toxicological effects (Gelderblom *et al.*, 2004).

IARC has concluded that there is inadequate evidence in humans for the carcinogenicity of toxins derived from *F. verticillioides* (as *F. moniliforme*), leading to a classification of Group 2B, possibly carcinogenic to humans (IARC, 1993). FB₁ was also classified as Group 2B (IARC, 2002) (Table 6.1).

3.4 Neural tube defects

Animal studies have demonstrated that fumonisin exposure can cause neural tube defects, possibly through the disruption of sphingolipid biosynthesis and consequent depletion of sphingolipids, which are critical for lipid raft functions, specifically folate processing via

the high-affinity folate transporter (Stevens and Tang, 1997; Sadler *et al.*, 2002; Gelineau-van Waes *et al.*, 2005). Neural tube defects are known to be associated with reduced folate levels, and cell membrane disruption induced by fumonisins could lead to reduced folate absorption through damage to the folate receptors on the membrane (Marasas *et al.*, 2004). More recently, elevation in sphingoid base 1-phosphates induced by fumonisins has been implicated in the induction of neural tube defects in mice (Gelineau-van Waes *et al.*, 2009).

A possible link between human neural tube defects and fumonisin consumption was suggested when a high rate of neural tube defects was recorded in babies of Mexican American women living in Texas who conceived during 1990–1991 (Hendricks, 1999), soon after the outbreaks of equine leukoencephalomalacia and porcine pulmonary oedema that occurred in 1989–1990 in the USA (Ross *et al.*, 1991). In the border region of Texas, exposure to fumonisins may be elevated due to frequent consumption of contaminated maize.

In a case–control study in this region of Texas (Missmer *et al.*, 2006), moderate tortilla consumption in the first trimester of pregnancy was associated with an increased risk of neural tube defects compared with low consumption. However, high consumption was not associated with increased risk. A similar result was found using estimates for fumonisin intake from tortillas, whereas an increased sphinganine-to-sphingosine ratio was associated with increased risk, apart from the highest category.

High incidence rates of neural tube defects have been recorded in rural areas of Mpumalanga Province, South Africa, and in the Umzimkulu district of the former Transkei region in Eastern Cape Province, South Africa

(Ncayiyana, 1986; Venter *et al.*, 1995); in Hebei Province, China (Moore *et al.*, 1997; Marasas *et al.*, 2004); and in Guatemala (Marasas *et al.*, 2004) and Mexico, all areas where large quantities of maize are consumed.

3.5 Occupational exposure

There are no reports of occupational exposure to fumonisin and adverse health effects.

4. Ochratoxin A

Human exposure to ochratoxin A (OTA) occurs principally in Europe and Canada, where it comes from eating foods made from barley and wheat in which *Penicillium verrucosum* has grown. Minor sources include meat, especially pork, from animals fed contaminated grain. In tropical and subtropical countries, OTA consumption is much lower, resulting from contamination due to growth of *Aspergillus carbonarius* and, less commonly, *A. niger* in coffee, cocoa and cocoa products, and dried fruit, and sometimes in cereals, including sorghum, maize, and millet (see Chapter 1).

OTA has been the subject of an Environmental Health Criteria document (WHO, 1990) and JECFA evaluations (WHO, 1991, 2001, 2002, 2007, 2008). JECFA noted that neither a conclusive association between OTA intake and human cancer nor the mechanism by which OTA is carcinogenic in animals has been established. JECFA has confirmed a provisional tolerable weekly intake (PTWI) for OTA of 100 ng/kg bw/week (WHO, 2001, 2008). It is noteworthy that risk assessment indicated that acute toxicity of OTA occurred in animals at lower levels than did long-term effects such as carcinogenicity, so this PTWI is based on acute toxicity.

4.1 Mechanisms

Recent reviews have extensively summarized evidence on the absorption, distribution, metabolism, and mechanisms of action of OTA (Pfohl-Leschkowicz and Manderville, 2007; Marin-Kuan *et al.*, 2008; Mally and Dekant, 2009). Wide species differences have been reported in the serum half-life of OTA *in vivo*. In humans, the elimination of OTA follows a two-phase pattern, a fast excretion followed by a slow clearing, with a calculated plasma half-life of 35 days. Even infrequent exposure (consumption of contaminated food once a week or even once a month) can result in persistent blood levels of OTA (Studer-Rohr *et al.*, 2000). Blood samples from healthy people in European countries show OTA levels of 0.1–40 ng/mL (WHO, 2008). The parent molecule is the major compound found in blood, whereas ochratoxin α is the major component detected in urine (Studer-Rohr *et al.*, 2000).

OTA is absorbed from the gastrointestinal tract in mammals and becomes strongly bound to plasma proteins (predominantly albumin) in blood, whereby it is distributed to the kidneys, with lower concentrations in liver, muscle, and fat. OTA is metabolized by several different CYP enzymes, depending on the species and tissue involved. In cells expressing human CYP enzymes, the main metabolite was 4(*R*)-hydroxy-OTA formed by CYP1A2, 2B6, 2C9, 2D6, and 2A6, whereas the 4(*S*)-hydroxy-OTA derivative was formed by only CYP2D6 and 2B6 (Pfohl-Leschkowicz and Manderville, 2007). Identified OTA metabolites include not only these two hydroxylated species but also 10-hydroxy-OTA and ochratoxin α , which is formed by hydrolysis of the peptide bond in OTA and thus lacks the phenylalanine moiety and consequently is non-toxic. OTA may also be metabolized

Table 6.1. IARC Monographs evaluations of carcinogenic hazards of mycotoxins to humans

| Mycotoxin | Monographs volume (year) | Degree of evidence of carcinogenicity | | Overall evaluation of carcinogenicity to humans ^a |
|--|--------------------------|---------------------------------------|-------------------------|--|
| | | In humans | In animals | |
| Aflatoxins, naturally occurring mixtures of | 56 (1993), 82 (2002) | Sufficient | Sufficient | Group 1 |
| Aflatoxin B ₁ | 56 (1993) | Sufficient | Sufficient | |
| Aflatoxin B ₂ | 56 (1993) | | Limited | |
| Aflatoxin G ₁ | 56 (1993) | | Sufficient | |
| Aflatoxin G ₂ | 56 (1993) | | Inadequate | |
| Aflatoxin M ₁ | 56 (1993) | Inadequate | Sufficient | Group 2B |
| Toxins derived from <i>Fusarium verticillioides</i> ^b | 56 (1993) | Inadequate | Sufficient | Group 2B |
| Fumonisin B ₁ | 82 (2002) | Inadequate | Sufficient | Group 2B |
| Fumonisin B ₂ | 56 (1993) | | Inadequate | |
| Fusarin C | 56 (1993) | | Limited ^c | |
| Ochratoxin A | 56 (1993) | Inadequate | Sufficient | Group 2B |
| Toxins derived from <i>Fusarium graminearum</i> , <i>F. culmorum</i> , and <i>F. crookwellense</i> | 56 (1993) | Inadequate | | Group 3 |
| Deoxynivalenol | 56 (1993) | | Inadequate ^d | |
| Nivalenol | 56 (1993) | | Inadequate | |
| Zearalenone | 56 (1993) | | Limited ^e | |
| Citrinin | 40 (1986) | Inadequate | Limited ^f | Group 3 |
| Patulin | 40 (1986) | Inadequate | Inadequate | Group 3 |
| Toxins derived from <i>Fusarium sporotrichioides</i> | 56 (1993) | Inadequate (no data) | | Group 3 |
| T-2 toxin | 56 (1993) | | Limited ^g | |

^a Group 1, carcinogenic to humans; Group 2A, probably carcinogenic to humans; Group 2B, possibly carcinogenic to humans; Group 3, not classifiable as to its carcinogenicity to humans; Group 4, probably not carcinogenic to humans.

^b Formerly known as *Fusarium moniliforme*. Fumonisin B₁ is also produced by additional *Fusarium* species.

^c Fusarin C caused marginal increases in incidences of papillomas and carcinomas of the oesophagus and forestomach when given by gavage to mice and rats.

^d More recent carcinogenicity studies of deoxynivalenol in mice have been negative (Iverson *et al.*, 1995; Lambert *et al.*, 1995) and reinforce the conclusion of the 1993 IARC Working Group that there is inadequate evidence in experimental animals for the carcinogenicity of deoxynivalenol.

^e Zearalenone caused increased incidences of hepatocellular and pituitary tumours in mice of both sexes when given in the diet, but no carcinogenic effect was seen in rats.

^f Citrinin caused benign epithelial tumours (clear cell adenomas) of the kidney in male rats in one experiment. Whereas this study (Arai and Hibino, 1983) was unequivocally positive, other studies in other strains of rat at lower doses were negative (IARC, 1986). A single positive study with only benign tumours as findings is considered limited evidence of carcinogenicity by the IARC Monographs criteria.

^g T-2 toxin caused increased incidences of liver cell tumours and lung tumours in male mice when given in the diet.

by cyclo-oxygenase, lipoxygenase, and epoxygenase, particularly in extrahepatic organs such as the kidney, to yield reactive oxygen species, which in turn may result in oxidative damage.

OTA competitively inhibits phenylalanine-tRNA ligase, resulting in inhibition of protein synthesis as well as RNA and DNA synthesis. In animals, acute toxic effects of OTA can be inhibited by co-administration of phenylalanine (FAO/WHO/UNEP, 1999).

The formation of DNA adducts by OTA and their potential role in cancer induction has been investigated (Mally and Dekant, 2009; Mantle *et al.*, 2010), and hypotheses for the formation include direct adduct formation induced by OTA after metabolic activation via an OTA phenoxy radical and indirect DNA damage resulting from oxygen radical formation as mentioned above (Pfohl-Leszkowicz and Manderville, 2007).

The concentration of OTA-specific transporters in tissues has been proposed to explain the relative species, sex, and target organ sensitivities to OTA toxicity (reviewed in Dietrich *et al.*, 2005). Another contributor to selective sensitivity is the extent of albumin binding, which markedly decreases the uptake of OTA by transporters (Bow *et al.*, 2006). Mechanisms that account for the toxicity and carcinogenicity of OTA without invoking the production of OTA DNA adducts have also been proposed (reviewed in Mally and Dekant, 2009) and typically involve alterations in expression of genes regulating rates of cell proliferation and cell death. Potential biomarkers of effect in target tissues include the development of unique gene expression profiles specific to alterations of gene expression induced by OTA. Genes include those involved in cellular defence, cell proliferation, and cell death

(reviewed in Marin-Kuan *et al.*, 2008; Adler *et al.*, 2009; Mally and Dekant, 2009) and in oxidative stress (Arbillaga *et al.*, 2008; Cavin *et al.*, 2009). Changes in urinary metabolite profiles, obtained using GC-MS, LC-MS, and [³H]-NMR, have been used as a metabonomic approach to assess the potential of these changes for development of a predictive model for OTA toxicity (Sieber *et al.*, 2009). Although the results were not specific for OTA, they were indicative of kidney damage and general toxicity, and this approach could prove to be of value for discovery of more specific mechanisms unique to OTA.

4.2 Acute toxicity

The kidney is the major target organ for adverse effects from OTA (Pfohl-Leszkowicz and Manderville, 2007; WHO, 2002). Short-term toxicity studies in mice, rats, dogs, and pigs have shown both time- and dose-dependent development of progressive nephropathy. Significant sex and species differences exist, as well as differences due to route of administration. Other toxic effects include cardiac and hepatic lesions in rats, lesions of the gastrointestinal tract and lymphoid tissues in hamsters, myelotoxicity in mice, and kidney lesions in chickens. Pigs appear to be the most sensitive species to the nephrotoxic effects; the lowest-observed-effect level (8 µg/kg bw) was used as the basis for establishing the PTWI.

A great deal of work has been undertaken recently to elucidate the likely mechanisms of toxicity. Degenerative changes in the proximal tubules of the kidney have been the most common effects seen in animal species studied. However, it has not been possible to determine just what acute effects, if any, OTA has on humans (WHO, 2008).

4.3 Cancer

Evidence for the carcinogenicity of OTA is principally from studies in experimental animals. OTA is carcinogenic to laboratory rats and mice, causing HCC in mice and kidney carcinomas in mice and rats. The mechanism of carcinogenic action has not been firmly established.

Reports of increased cancer risk in humans who consumed OTA have been limited to descriptive studies. No analytical epidemiological studies were available to IARC at the time of the evaluation of OTA (IARC, 1993). The descriptive studies generally focused on the co-occurrence of Balkan endemic nephropathy, a fatal chronic renal disease, and higher than expected rates of urinary tract tumours, including tumours of the kidney and urinary bladder, in Bulgaria and other Balkan countries. Available studies have not established that this increased cancer incidence was due to exposure to OTA.

IARC has concluded that there is sufficient evidence that OTA is carcinogenic in experimental animals but inadequate evidence that OTA increases cancer risk in humans. OTA has therefore been classified as Group 2B, possibly carcinogenic to humans (IARC, 1993) (Table 6.1).

4.4 Occupational exposure

OTA is found in the spores of *P. verrucosum* on grains and also *A. carbonarius* as well as some strains of *A. niger* on grapes and coffee beans. Some occupational studies in Europe reported elevated OTA levels in plasma in workers exposed to grain dust (Pfohl-Leszkowicz and Manderville, 2007). From what was inferred to be a massive exposure to OTA from working in a confined space with grain contaminated by

A. ochraceus, a farmer developed acute renal disease after temporary respiratory distress (Di Paolo *et al.*, 1993). Only limited estimates of inhalation exposure to OTA from occupational exposure are available (Mayer *et al.*, 2007).

5. Deoxynivalenol

Deoxynivalenol (DON) is produced in cereals, especially wheat, as the result of growth of *Fusarium graminearum* and related species (see Chapter 1). JECFA has established a PMTDI for DON of 1 µg/kg bw/day on the basis of the NOEL for body weight reduction in mice in a 2-year bioassay and a safety factor of 100 (WHO, 2001, 2011). The NOEL in mice was 100 µg/kg bw/day (Iverson *et al.*, 1995).

5.1 Mechanisms

DON is directly toxic via an epoxide moiety and thus does not require metabolic activation to exert its biological effects. Low-level trichothecene exposure in animal models has been shown to modulate the expression of several cytokines and chemokines that are key regulators of immune function (Pestka, 2008). Exposure to DON causes the upregulation of the mRNAs responsible for production of cytokines, chemokines, and other immune-related proteins and can also induce gene transcription. In addition, DON modulates numerous physiological processes controlled by mitogen-activated protein kinases (MAPKs). These include processes controlling cell growth, differentiation, and apoptosis, which are all crucial for signal transduction in the immune response (Pestka, 2008). Thus, in addition to altered cytokine expression, alterations in MAPK expression are likely to also contribute to the immune dysregulation and toxicity of DON and other trichothecenes. Also associated

with MAPK activation by DON is the activation of processes leading to the ribotoxic stress response, which is induced by other translational inhibitors that, like DON, bind to or damage a specific region at the 3' end of the 28S rRNA. The ribosome plays a key role in the ribotoxic stress response by serving as scaffolding for interactions between various MAPKs (Pestka, 2008).

DON toxicity studies have recently revealed several possible approaches for developing useful biomarkers of its effects. For example, DON exposure in mice results in upregulation of several suppressors of cytokine signalling. These suppressors are known to impair growth hormone signalling (Pass *et al.*, 2009). Impairment of the growth hormone axis precedes the growth retardation in the mouse induced by DON (Amuzie and Pestka, 2010). Oral DON perturbs the growth hormone axis by suppressing two growth-related proteins, IGFALS and IGF1. Thus, reduced expression of these two proteins in conjunction with elevated urinary DON levels could potentially serve as biomarkers of effect.

Detoxification varies by species; metabolism by gut microflora generates a de-epoxy metabolite (DOM-1), and conjugation to glucuronic acid is catalysed by UDP-glucuronyltransferase (reviewed by Pestka and Smolinski, 2005; Wu *et al.*, 2007). In humans, DON-glucuronide has been reported in urine samples in several studies (Turner *et al.*, 2008a, 2008b, 2008c). In contrast, information on DOM-1 is limited. An absence of de-epoxidase activity in a small series of human faecal samples was reported (Sundstøl-Eriksen and Pettersson, 2003), whereas, in apparent contrast, a study in France of farmers exposed to grain handling reported detection of DOM-1 in a proportion of subjects (Turner *et al.*, 2010).

5.2 Acute toxicity

In animals, DON has a wide range of proven toxicities, including feed refusal, decreased weight gain, gastroenteritis, cardiotoxicity, teratogenicity, and immunotoxicity (Rotter *et al.*, 1996; Meko *et al.*, 2001; Pestka *et al.*, 2004; Pestka and Smolinski, 2005; Gray and Pestka, 2007; Amuzie and Pestka, 2010).

DON can cause acute poisoning in humans, where severe gastrointestinal toxicity is the primary symptom. Consumption of cereals contaminated with DON has been associated with numerous poisoning incidents in China between 1961 and 1991 (see Pestka and Smolinski, 2005) and a major outbreak in India (Bhat *et al.*, 1989); in some of these episodes, tens of thousands of individuals were affected. In these outbreaks, symptoms were analogous to those observed in animals, notably a rapid onset, nausea, vomiting, abdominal pain, diarrhoea, headache, dizziness, and fever. In an episode in the Kashmir valley, DON levels in wheat ranged from 0.4 mg/kg to 8.4 mg/kg (Bhat *et al.*, 1989), and in China DON poisoning was linked to wheat contaminated at DON levels between 0.3 mg/kg and 100 mg/kg (Pestka and Smolinski, 2005). These data suggest that acute toxicity may occur at exposures estimated in the low µg/kg bw/day range.

5.3 Cancer

Minimal data are available on the carcinogenicity of DON in either humans or experimental animals.

A 2-year bioassay in B6C3F1 mice of both sexes fed DON in the diet at concentrations of 0, 1, 5, or 10 mg/kg showed no increase in the incidence of pre-neoplastic or neoplastic lesions in the liver or other tissues (Iverson *et al.*, 1995). DON was also tested for its ability to initiate or promote skin tumours when applied topically

to the skin of female SENCAR mice, with negative results (Lambert *et al.*, 1995). No studies have reported on the carcinogenicity of DON in humans. Oesophageal cancer in humans has been anecdotally linked to consumption of grains infected with *Fusarium* species that produce DON and other mycotoxins, but no analytical epidemiological studies link DON to the occurrence of any human cancer (IARC, 1993).

IARC has concluded that there is inadequate evidence in both humans and experimental animals for the carcinogenicity of DON. DON and other toxins derived from *F. graminearum*, *F. culmorum*, and *F. crookwellense* have therefore been categorized as Group 3, not classifiable as to their carcinogenicity to humans (IARC, 1993) (Table 6.1).

5.4 Occupational exposure

Inhalation exposure to DON has been the subject of several health hazard evaluations. *Fusarium* head blight in wheat resulting from infection by *F. graminearum* or *F. culmorum* begins at the outside of the grain head and moves inward. As a result, most DON is found in the outer layers of the kernel and the chaff (Miller, 1994b; Snijders, 1994). Grain dusts can contain quite high concentrations of DON and sometimes of other mycotoxins, which are not always present in the kernels. For example, other fungi, including *F. sporotrichioides*, can grow on what becomes the chaff so that small amounts of T-2 toxin can be present.

Air samples collected in grain elevators in Canada contained a mean airborne concentration of 37 ng/m³ DON and a maximum of 2.59 µg/m³ DON. Airborne dust from the same source contained 0.5–5.8 mg/kg DON, 1 mg/kg T-2 toxin, and low levels of HT-2 toxin (De Mers, 1994). Concentrations of dusts, fungal spores, and DON associated with

handling of grain on farms in Finland, including grain drying, milling, and cattle feeding, were similar to those reported from Canada (Lappalainen *et al.*, 1996).

Studies of DON concentrations during grain handling in Germany reported a median concentration of 2 ng/m³ and a maximum of 703 ng/m³ (Mayer *et al.*, 2007). In France, urinary biomarkers for DON (DON, DOM-1) were higher in active farmers, particularly those from larger farms, than in retired farmers, whose exposure was from diet (Turner *et al.*, 2010).

Epidemiological studies have been conducted in Norway relating to occupational exposures of male and female farmers to mycotoxins. Norwegian grains (wheat, oats, and barley) are affected mainly by *Fusarium* head blight, and various trichothecenes and culmorins have been reported as common (Langseth and Elen, 1996; Ghebremeskel and Langseth, 2001). A longitudinal survey of farmers over more than two decades suggested a relationship between grain farming and mid-pregnancy deliveries in the families of farmers, possibly linked to mycotoxins (Kristensen *et al.*, 1997). Small increased relative risks were observed in several cancers in female but not male farmers (Kristensen *et al.*, 2000).

6. Zearalenone

Zearalenone (ZEA) is a non-steroidal estrogenic mycotoxin. It is produced principally by *F. graminearum* and related species, and consequently occurs wherever DON occurs, most notably as a contaminant of maize, wheat, barley, oats, rye, sorghum, millet, and rice. Distribution of ZEA is worldwide (Zinedine *et al.*, 2007). Estimates of human exposure from dietary sources are generally in the range of 1–30 ng/kg bw/day (Zinedine *et al.*, 2007). JECFA has established

a PMTDI for ZEA of 0.5 µg/kg bw/day based on the NOEL for hormonal effects in pigs (WHO, 2001).

6.1 Mechanisms

ZEA is metabolized during absorption by the intestinal tissue in pigs. Metabolism involves reduction of the 6-keto group of ZEA and results in formation of α- and β-zearalenol as well as, upon further reduction, α- and β-zearalanol, all of which can be conjugated in turn to glucuronic acid (WHO, 2001). Few data are available in relation to metabolism in humans. Studies of liver microsomes in vitro have suggested a high rate of α-zearalenol production compared with that of β-zearalenol in pigs and humans, a point of importance because of the greater relative estrogenicity of α-zearalenol compared with ZEA (Fink-Gremmels and Malekinejad, 2007). In other words, the formation of α-zearalenol may be considered a bioactivation step contributing to the estrogenic effects of ZEA.

ZEA and its metabolites can bind to estrogen receptors, resulting in various changes consequent to binding to elements in the nucleus responsive to estrogens. In addition, however, ZEA is a competitive substrate for enzymes involved in steroid synthesis and metabolism and therefore has the potential to act as an endocrine disruptor. ZEA can activate the pregnane X receptor by displacement of a co-repressor and recruitment of co-activators (Ding *et al.*, 2006). Thus, ZEA could have widespread effects on gene expression as a result of the modified activity of this nuclear transcription factor.

6.2 Acute toxicity

ZEA is considered to be of relatively low acute toxicity. No reports have appeared of acute poisoning due to ZEA in humans.

6.3 Cancer

ZEA resulted in an increased incidence of liver cell and pituitary tumours in mice, consistent with a hormonal mode of carcinogenic action (IARC, 1993). No carcinogenic effect was seen in rats, however, and overall animal carcinogenicity data for ZEA were considered limited (IARC, 1993). No studies of human carcinogenicity have been reported for ZEA (Table 6.1).

ZEA was measured in endometrial tissue from a small group of women with endometrial adenocarcinomas, endometrial hyperplasia, or normal proliferative endometria (Tomaszewski *et al.*, 1998). ZEA in blood samples has also been investigated in some small studies of early onset of puberty in Hungary (Szuets *et al.*, 1997) and in Italy (Massart *et al.*, 2008). In the study in Italy, there was a suggestion that elevated serum ZEA and α -zearalenol levels were associated with early puberty in 6 of the 17 girls examined from a rural area, but no positive samples were seen in the 15 patients from an urban area.

6.4 Occupational exposure

No studies of occupational exposure of humans have been reported for ZEA.

7. Occupational exposures to grain and groundnut dusts

Grain dusts present an occupational hazard when protection of workers is inadequate, and several health consequences are possible, including the allergic disease hypersensitivity pneumonitis, endotoxiosis, and organic dust toxic syndrome (Rylander and Jacobs, 1994; Sorenson and Lewis, 1996). Endotoxiosis is a result of exposure to bacterial endotoxin, and organic dust toxic syndrome (also called pulmonary mycotoxicosis, toxic organic dust syndrome, or grain fever), as far as is known, is not caused

by mycotoxins. Inhalation of silica, allergens, mycotoxins, and triple-helical glucans contained in airborne dusts is a potential health risk, rarely from systemic exposure but from effects on lung biology. Outdoor work is normally not a problem, except when handling the most damaged maize, small grains, and groundnuts. In contrast, handling contaminated grains, especially damaged grains, in any confined space (e.g. grain storage, storage bin cleaning, animal feeding in barns, indoors) carries a more significant risk.

The allergic disease hypersensitivity pneumonitis, also called extrinsic allergic alveolitis or farmer's lung, develops through repeated exposure to allergens. Dust from any mouldy crop, such as straw, maize, small grains, or groundnuts, can cause the condition. Symptoms may include shortness of breath, a dry cough, a sudden general feeling of sickness, fevers and chills, a rapid heart rate, and rapid breathing. The symptoms are serious, and once an allergic reaction begins, the person will always have the potential for symptoms with exposure to the offending fungi. Long-term exposure can cause permanent lung damage, physical disability, or even death (Sorenson and Lewis, 1996; Schenker *et al.*, 1998; Girard *et al.*, 2009).

The reproductive structures of many fungi are known to contain mycotoxins or low-molecular-weight toxic compounds, often in high concentrations (Sorenson, 1999). Most is known about the fungi that produce toxins important in agriculture. The conidia of *A. flavus*, *A. parasiticus*, *F. graminearum*, and *F. sporotrichioides* contain very high concentrations of toxins, particularly in the case of the species that produce aflatoxins. The spores of the two *Aspergillus* species have been reported to contain 100–1100 μg aflatoxin/g, or approximately 10^{-4} moles (Wicklow and

Shotwell, 1983). Several interesting toxins have been found in sclerotia of various *Aspergillus* species, again at higher concentrations than occur either in culture or in affected crops (Gloer *et al.*, 1988; Wicklow *et al.*, 1988), some of which are thought to be present in conidia along with kojic acid and other *A. flavus* toxins. Spores of *F. graminearum* contained 30 $\mu\text{g/g}$ of T-2 toxin and those of *F. sporotrichioides* 50 $\mu\text{g/g}$ (both approximately 10^{-5} moles). The spores of many species of toxigenic fungi have been demonstrated to contain mixtures of the toxins associated with the species.

The high-molecular-weight toxic compound present in spores and in spore and mycelial fragments from the anamorphic *Trichocomaceae* (i.e. *Penicillium*, *Aspergillus*, and related hyphomycetes) is (1 \rightarrow 3)- β -D-glucan in the triple-helical form (Rand *et al.*, 2010). In the species tested so far, the concentration is 1–11 pg/spore (Foto *et al.*, 2004; lossifova *et al.*, 2008).

Inhalation of intact spores leads to little net exposure because their relatively large size enables entrapment and removal by lung defence mechanisms. However, in outdoor air, exposure is primarily to spore and mycelial fragments (Green *et al.*, 2012) and to dusts (particulate matter < 2.5 μm in diameter), which efficiently penetrate deep into the lung (Buczaj, 2008; see Miller *et al.*, 2010). Inhalation of toxins affects macrophage function and other aspects of lung biology. Pure compounds tested for effects in macrophages include fumonisin, aflatoxin (Liu *et al.*, 2002), T-2 toxin (Sorenson *et al.*, 1986), and the aflatoxin precursor sterigmatocystin (Miller *et al.*, 2010). DON has been tested for effects in peritoneal macrophages (Ayrat *et al.*, 1992).

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